

**Distribution of *Alexandrium fundyense* (dinophyceae) cysts in Greenland and
Iceland, with an emphasis on viability and growth in the Arctic**

Mindy L. Richlen^{1*}, Oliver Zielinski², Lars Holinde², Urban Tillmann³, Allan
Cembella³, Yihua Lyu^{1,4}, Donald M. Anderson¹

¹ Woods Hole Oceanographic Institution, 266 Woods Hole Rd., MS 32, Woods Hole,
Massachusetts 02543, USA

² Institute for Chemistry and Biology of the Marine Environment, University of
Oldenburg, 26111 Oldenburg, Germany

³ Alfred-Wegener-Institut, Helmholtz Zentrum für Polar-und Meeresforschung, Am
Handelshafen 12, 27570 Bremerhaven, Germany

⁴ State Key Lab of Marine Environmental Science, School of Life Science, Xiamen
University, Xiamen, 361102, China

*Corresponding author E-mail: mrichlen@whoi.edu

Running page head: *Alexandrium fundyense* cysts in Arctic waters

Key words: Arctic, *Alexandrium*, dinoflagellate, cysts, harmful algal bloom

ABSTRACT

The bloom-forming dinoflagellate *Alexandrium fundyense* has been extensively studied due its toxin-producing capabilities and consequent impacts to human health and economies. This study investigated the prevalence of resting cysts of *A. fundyense* in western Greenland and Iceland to assess the historical presence and magnitude of bloom populations in the region, and to characterize environmental conditions during summer, when bloom development may occur. Analysis of sediments collected from these locations showed that *Alexandrium* cysts were present at low to moderate densities in most areas surveyed, with highest densities observed in western Iceland. Additionally, laboratory experiments were conducted on clonal cultures established from isolated cysts or vegetative cells from Greenland, Iceland, and the Chukchi Sea (near Alaska) to examine the effects of photoperiod interval and irradiance levels on growth. Growth rates in response to the experimental treatments varied among isolates, but were generally highest under conditions that included both the shortest photoperiod interval (16h:8h light:dark) and higher irradiance levels ($\sim 146\text{-}366\ \mu\text{mol photons m}^{-2}\ \text{s}^{-1}$), followed by growth under an extended photoperiod interval and low irradiance level ($\sim 37\ \mu\text{mol photons m}^{-2}\ \text{s}^{-1}$). Based on field and laboratory data, we hypothesize that blooms in Greenland are primarily derived from advected *Alexandrium* populations, as low bottom temperatures and limited light availability would likely preclude in situ bloom development. In contrast, the bays and fjords in Iceland may provide more favorable habitat for germling cell survival and growth, and therefore may support indigenous, self-seeding blooms.

Introduction

The bloom-forming dinoflagellate genus *Alexandrium* Halim emend Balech (1995) has been extensively studied due to its toxigenicity, particularly those taxa comprising the “*tamarense* species complex”, which includes *A. acatenella*, *A. catenella*, *A. excavatum*, *A. fundyense*, *A. tamarense*, and several closely related species formerly assigned to *Protogonyaulax* Taylor. The human illness caused by the toxins produced by *Alexandrium* is known as Paralytic Shellfish Poisoning (PSP), which is widespread in temperate waters around the world.

One strategy for the success of this dinoflagellate across such a range of habitats is that the life cycle of many species in this genus includes a benthic cyst stage. This life cycle stage allows cells to enter dormancy during unfavorable temperature or nutrient conditions, and survive in sediments during temperature extremes (i.e., overwinter), with seasonal germination inoculating vegetative cells into the water column only during intervals when conditions are suitable for growth (Anderson, 1998). Population development is thus possible in more locations than would otherwise be the case if year-round persistence in the plankton were the only means for survival. The cyst is also critical in species dispersal, as cells transported to new locations by storms, currents, wildlife, or humans can colonize an area by depositing cysts that germinate in subsequent years.

Due to the widespread and serious impacts of these blooms, the distribution, life cycle, taxonomy, and physiology of *Alexandrium tamarense* species complex taxa are relatively well-studied compared with many other globally distributed phytoplankton. Recent morphological, molecular, and mating studies indicate that the strains comprising the Group 1 (formerly North American) clade (Lilly et al. 2007) of this complex comprise a single species – *Alexandrium fundyense* (John et al. 2014a).

74 There is general agreement that the *tamarensis* complex should be split into separate
75 species, but there is some disagreement on the name to be used for that clade, with
76 some arguing that it should be *A. catenella* and not *A. fundyensis* (Fraga et al. 2015).
77 Here we use the name *A. fundyensis*, but recognize that this issue is not yet fully
78 resolved.

79 Although *A. fundyensis* is not considered to be endemic to the Arctic, several
80 recent reports of *Alexandrium* cysts, cells, and toxins from Arctic waters suggest that
81 suitable habitat for growth and bloom formation is present in this region. *Alexandrium*
82 *fundyensis* has been reported from coastal waters near Barrow, AK (Okolodkov 2005),
83 and recent work by several groups documented *A. fundyensis* cysts, vegetative cells,
84 and toxins in the Chukchi and Beaufort Seas (Gu et al. 2013, Natsuike et al. 2013).
85 Notably, the extraordinarily high densities of *A. fundyensis* cysts (maximum 10,600
86 cysts cm⁻³) observed in surface sediments from the Chukchi Sea are among the
87 highest ever reported for this species (Natsuike et al. 2013).

88 In waters east of North America and north of Europe, *A. fundyensis* has been
89 observed in plankton samples from the Labrador, Greenland, and Norwegian Seas
90 (Scholin 1998, Okolodkov 2005, Baggesen et al. 2012, Tillmann et al. 2016), and in
91 the Northwest Passage in the Canadian archipelago (M. Levasseur pers. comm.). PSP
92 toxins in blue mussels were recently reported for the first time from Iceland, along
93 with record high numbers of toxic *Alexandrium* spp. (>16,000 cells L⁻¹) (Burrell et al.
94 2013). Additionally, PSP toxins were detected in scallops at levels exceeding
95 regulatory limits for the first time in western Greenland (Baggesen et al. 2012), and *A.*
96 *fundyensis* was isolated and identified from nearby waters. Although it is clear that
97 environmental conditions in at least some areas of the Arctic foster cell growth and
98 bloom development, it is yet unclear whether and where the establishment of endemic

populations (via cyst germination) might be possible, and how climate-driven increases in bottom temperatures might influence the future range and magnitude of blooms in the region.

The aforementioned reports prompted the current investigation, which sought to better characterize the present distribution of this species in western Greenland and Iceland relative to environmental conditions, and to examine growth responses of Arctic isolates compared with those from temperate regions. We examined sediment samples collected in western Greenland and Iceland for *Alexandrium* cyst accumulations, an approach for assessing the presence and magnitude of bloom populations present in previous years, to better understand the prevalence of *Alexandrium* in Arctic waters. Additionally, we characterized the particle size distribution of sediment samples (sediment structure) to assess whether certain areas might favor the accumulation of higher cyst densities.

Data collected on the underwater light field, temperature, and sampling depth were used to infer the potential for germling survival and cell growth. Associated laboratory experiments were carried out with *A. fundyense* cultured isolates established during these surveys to examine their growth responses to the particular light intensities and photoperiod intervals that bloom populations would experience during summer months in the Arctic. Our goals were to: (1) provide a preliminary characterization of cyst densities in western Greenland and Iceland, including comparisons between fjord and external coastal habitats, (2) assess environmental factors (temperature, light, and photoperiod interval) that determine viability and growth of germling cells, (3) identify areas that might favor in situ bloom initiation, and (4) use these data to generate hypotheses regarding the origins and fate of PSP-toxin producing *Alexandrium* populations in this region. Many dinoflagellate species

form cysts, and thus studies like the one presented here provide information about how this adaptive strategy might influence the distributions of many species in a warming climate.

Materials and Methods

Field collections

Sediment sampling and data collection were carried out during a research oceanographic cruise aboard the *RV Maria S. Merian* (July 27-Aug 8, 2012); see also Cembella et al. (2016). This particular cruise leg (field campaign MSM 21/leg 3) included a comparative study of the west coasts and fjords of Greenland and Iceland (Fig. 1). Sediments were collected from a total of 20 stations to characterize the prevalence of *A. fundyense* cysts in western Greenland and Iceland. Samples in Greenland were collected from Uummannaq Fjord, the Vaigat, and Disko Bay, and from two stations near Cape Farewell, and from Arnarfjörður and Breiðafjörður in Iceland (Fig. 1). Sediments were collected with two Van Veen grab-samplers, each of which can extract sediments up to 20 cm deep, with a sampling area of 0.04 m² (small sampler) or 0.1 m² (large sampler). Samples for cyst enumeration, culture establishment and analysis of sediment characteristics were collected from the sediment surface layer (upper 10 cm) in the grab, and stored in anoxic conditions in the dark and at 2 °C until further processing (Anderson et al. 1987). Subsequent laboratory analyses were carried out at the Woods Hole Oceanographic Institution, Woods Hole, MA, USA.

For cyst enumeration, a homogenized 5 cm³ sediment sample was removed from each sample, resuspended with filtered seawater, sonicated with a Branson

Sonifier 250D at a constant 40-watt output for one minute, and sieved to yield a clean, 20-100 μm size fraction (Anderson et al. 2003). Cysts were then concentrated using a single density layer (1.4 g cm^{-3}) of NALCO 1060 colloidal silica (Nalco Company, IL, USA) (Schwinghamer et al. 1991, Anderson et al. 2003) following procedures described in Vahtera et al. (2014). *Alexandrium fundyense* cysts were stained with primulin (MP Biomedicals, LLC, OH, USA) and enumerated in each sample as described in Anderson et al. (2003).

Analysis of sediment structure

Sediment samples were collected at defined stations and frozen at -25°C until analysis at the University of Oldenburg, Oldenburg, Germany. For these analyses, the particle size distribution (PSD) from subsamples was determined using a laser scattering particle size analyzer (Horiba LA-950, Japan). To remove coarse fragments before measurement, subsamples were sieved through a 2 mm mesh sieve and treated with sodium meta-phosphate (NaPO_3 , 2% in water) due to presence of aggregates in the sample. The filtrate was examined with the laser particle size analyzer, which has a measurement range of $0.01 - 3000 \mu\text{m}$, providing relative composition of seven granulometric fractions from $<2 \mu\text{m}$ to $630 - 2000 \mu\text{m}$.

Water column properties and optical measurements

At each sampling location data on water column properties were collected with a CTD-rosette sampler, and above-water and in-water hyperspectral radiometric measurements were collected to investigate the optical properties of water masses (see also Garaba & Zielinski 2013, Holinde & Zielinski 2015). The CTD casts were

performed with a Seabird “sbe911+” CTD probe with sampling rosette at each station as a start-up to determine further key discrete sampling depths, e.g., to locate chlorophyll maxima. Live data acquisition was carried out via CTD-client onboard and data post-processing with Seasoftware V2 (Seabird, WA, USA). Salinity and depth were calculated from pressure values (UNESCO 1983), and temperature was corrected to ITS-90 (Preston-Thomas 1990). All CTD data are available from the WDC-Mare database system Pangaea® [doi:10.1594/PANGAEA.819731](https://doi.org/10.1594/PANGAEA.819731).

A HyperPro II profiling system (Satlantic, Halifax, Canada) was used to acquire bio-optical data for inherent and apparent optical properties (Hölinde & Zielinski 2015). The profiler consisted of one hyperspectral irradiance and one hyperspectral radiance sensor. A second hyperspectral irradiance sensor was mounted at an unshaded elevated position on the research vessel for reference measurements (E_s). On the profiler, the irradiance and radiance sensors measured downwelling (E_d) and upwelling (L_u) light, respectively.

Profiler measurements were conducted at selected stations depending on sea, weather, and daylight conditions. At these stations, three casts were typically performed. For each cast, the profiler was lowered until the downwelling light values were of the same order of magnitude as the background noise level of the sensor. Hyperspectral $E_d(\lambda)$ data were then processed with ProSoft 7.7.16 (Satlantic) and binned to 0.2 m depth intervals to calculate photosynthetically active radiation (PAR)

$$PAR(z) = \int_{400}^{700} (\lambda/hc) E_d(\lambda) d\lambda$$

where z is the depth in meters, λ is the wavelength in nanometers, h is Planck's constant and c is the speed of light. Additionally, the percentage of PAR reaching depth z with reference to $PAR(0^+)$ calculated from $E_s(\lambda)$ was determined according to:

$$\%P(z) = PAR(z)/PAR(0^+) \times 100$$

Based on $\%PAR(z)$, the 1% depth of PAR, a common indicator for the depth of the euphotic zone, was derived together with the maximum wavelength present at that depth. The mean values of the available profiles were used for all calculated profiler data.

Irradiance and photoperiod experiments

A subset of plankton and sediment samples was also used to establish *Alexandrium* spp. cultures (Supplementary Table S1), either from single cell isolations from plankton samples (Tillmann et al. 2016), or from germinated cysts. Additionally, isolates were established from cysts in sediments collected from the Chukchi Sea, which were kindly provided by Dr. Haifeng Gu (Third Institute of Oceanography, Xiamen, P.R. China). Sediment samples were processed as described above, and cysts were isolated via micropipetting and placed in individual wells of 24-well tissue culture plates containing f/2(-Si) growth medium (Guillard & Ryther 1962). Plates were incubated for approximately one week at 10 °C under a 14h:10h light:dark photoperiod cycle. Wells were examined daily for germination and once sufficient motile cells were observed, individual cells were isolated, washed, and placed singly into tubes containing f/2(-Si) medium. Cultures were initially maintained at 10 °C, but were subsequently maintained at 15 °C due to improved growth at the higher temperature. Species designations of isolates were determined by

sequencing the highly variable D1-D2 domains of the large subunit ribosomal RNA gene (LSU rRNA) (Tillmann et al. 2016; D.M. Anderson unpub. data).

A series of laboratory experiments were performed to assess the effects of irradiance and photoperiod interval on growth responses of *A. fundyense* under the particular light conditions that bloom populations would experience during summer in the Arctic. These experiments were carried out with three isolates each from Greenland, Iceland, and the Chukchi Sea; three isolates originating from a temperate location, the Gulf of Maine (GOM), were also examined (Supplementary Table S1). The Greenland isolate E516 died before the experiments could be completed; it was therefore necessary to use a different isolate (P3H8) in the 24h:0h light:dark (L:D) photoperiod treatment (see below). Experiments were performed in an incubator at a constant 12 °C; each isolate was grown in triplicate under irradiance levels of 37, 92, 146, 183, 275, and 366 $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$, which were established on four shelves by combining different light settings on each shelf with nylon window screen (1-2 layers) to provide additional shading. The lowest irradiance level was selected based on prior studies of *A. fundyense* from the Northwest Atlantic (Etheridge & Roesler 2005), in which growth rates in response to irradiance were lowest under 25 and 50 $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$ (but at 20 °C, higher than the temperature level used in these experiments).

Irradiance received by the cultures was measured with a digital scalar irradiance meter (Model QSP-170, Biospherical Instruments, CA, USA) equipped with a probe QSL-100. Experiments were replicated under three different photoperiod intervals: 24h:0h, 20h:4h, and 16h:8h L:D. Preliminary studies confirmed a linear correlation between in vivo fluorescence and cell concentrations, and population growth was subsequently monitored by in vivo fluorescence measured with a 10-AU

fluorometer (Turner Designs, USA) in a 25 mm cuvette. Fluorescence was measured in each tube at the same time (~10:30) three times per week, and tubes were shaken by hand to distribute the cells uniformly in the medium before measuring fluorescence. Growth data were collected from three technical replicates of each isolate.

The intrinsic growth rate was calculated over the exponential phase of growth (as inferred from a semi-log plot of fluorescence versus time; see Guillard 1973, Wood et al. 2005) by the following equation:

$$\mu = \frac{\ln(N_1/N_0)}{t_1 - t_0}$$

in which μ (day^{-1}) represents the growth rate, and N_1 and N_0 represent the fluorescence at times t_1 and t_0 , respectively.

Statistical analyses

Statistical analyses to examine the effects of irradiance and photoperiod interval on growth were performed with JMP 11 software (SAS Institute, NC, USA). Datasets were first grouped by region (Greenland, Iceland, Chukchi, GOM) for these analyses. Effects of irradiance on growth were compared among regions, but within each photoperiod interval, and growth response data for each photoperiod interval (at all irradiance levels) were also pooled and compared. Growth data were not normally distributed, and it was not possible to achieve normality by transforming the data; non-parametric Welch's ANOVA and Wilcoxon rank sum tests were used instead for these comparisons, with $\alpha = 0.5$.

Additionally, principal components analysis (PCA) was performed with Primer v6.0 (Primer-E, Plymouth, UK) to examine correlations between cyst abundance, components of the sediment structure, and depth, and to evaluate regional clustering among samples.

Results

Temperature profiles and bio-optical parameters

Clear differences in temperature profiles were observed among the sampling locations (Fig. 2). Temperature measured in profiles from Uummannaq Fjord (transect distance 0–200 km), the Vaigat (250–450 km), and Disko Bay (>450 km) generally ranged from ~0–7 °C, and values throughout much of the water column were <4 °C. Maximum temperatures of ~11–12 °C were only found in surface waters at two locations. Cold melt water from the Perlerfiup Sermia glacier at the head of the Perlerfiup Kangerlua, a tributary fjord of the Uummannaq Fjord system, and representing the starting point of the transect, was detected throughout section distance (0–150 km) between 20 and 200 m depth with temperatures <1 °C.

Water temperatures in Icelandic fjords were much higher, ranging from 2–12 °C in Arnarfjörður and 6–15 °C in Breiðafjörður (Fig. 2). With the exception of the deepest areas of Arnarfjörður (>60 m), water temperatures in this fjord system were generally >8 °C throughout much of the water column.

Analysis of light availability from radiometric profiles showed that the 1% PAR depth was <46 m at all stations surveyed, ranging from 15.9 m (St 535) to 45.7 m (St 522) (Fig. 1, Table 1). Maximum wavelength at the 1% level was shifted from below 500 nm for all Greenland stations to 530 ± 30 nm for Iceland stations due to

increased presence of colored dissolved organic matter (CDOM) absorbing ultraviolet and blue spectral components (data not shown).

Distribution and abundance of Alexandrium cysts

Alexandrium fundyense resting cysts were observed in sediments from all stations surveyed in Greenland and Iceland, with the exception of St 523 and 524, located near Cape Farewell, Greenland (Fig. 1). Cysts of several other dinoflagellate taxa (*A. minutum*, *A. ostenfeldii*, *Protoceratium* sp., *Protoperidinium* sp., *Scrippsiella* sp.) were also observed (but not quantified) in many of the samples, including those collected from Cape Farewell. Cyst concentrations in sediments collected from the other Greenland stations ranged from 2 to 37 cysts cm⁻³ (mean \pm SD: 9 ± 11 cysts cm⁻³), with highest concentrations found in Uummannaq Fjord (Fig. 3). Cyst concentrations in Iceland were higher, ranging from 15 to 408 cysts cm⁻³ (mean \pm SD: 109 ± 127 cysts cm⁻³). In Iceland, highest cyst concentrations were observed at St 538 (408 cysts cm⁻³) in Breiðafjörður, followed by St 529 (124 cysts cm⁻³) and St 537 (120 cysts cm⁻³) in Arnarfjörður and Breiðafjörður, respectively (Fig. 3). The sediment sampling regimes differed substantially between Greenland and Iceland stations with respect to sampling depth. In Greenland, samples were collected from depths ranging from 135 to 550 m, with the majority of samples collected at depths >200 m (Fig. 4). In Iceland, however, sampling depths ranged from 51 to 330 m, and all but one sample were collected from depths <200 m.

Sediment characterization

Seven granulometric fractions ranging from <2 to 2000 μ m were quantified in each of the sediment samples. The most apparent differences among samples were the

higher proportions of fine silt in samples from Greenland compared with those from Iceland (Fig. 5), and the higher proportions of coarser, sandy sediments (F_{63-200} , $F_{200-630}$) in samples from St 523 and 524, collected near Cape Farewell. With the exception of these two stations, finer sedimentary fractions ($F_{<63}$) comprised 50% or more of the particle size fractions of each sample (Fig. 5).

In the principal component analysis (PCA) based upon data on sediment characteristics, water depth, and cyst abundance, the first two principal components accounted for 72.8% of the variance, with the third accounting for an additional 11.8%. The strongest correlations (positive or negative) for the first principal component were with $F_{2-6.3}$ and F_{63-200} . For the second principal component, the strongest correlations were with F_{63-200} and cysts cm^{-3} . In the PCA plot, clustering according to region was observed (Fig. 6). The first cluster comprised the samples collected from Iceland and St 516 (Disko Bay, Greenland), whereas the second comprised those from the Cape Farewell stations (St 523 and St 524), and the third comprised the remaining Greenland samples.

Photoperiod interval and irradiance experiments

Patterns of growth responses to the experimental treatments varied widely among isolates but were strain- rather than region-specific. In all cases, both photoperiod and irradiance had significant effects on growth, with the highest growth rate (0.28 day^{-1}) observed in these experiments for Iceland isolate D3 grown under constant light (24h:0h L:D) but at the lowest irradiance level (Supplementary Fig. S1). The next highest growth rates were observed for the same isolate grown under constant light, at the next two lowest irradiance levels (~ 92 and $\sim 147 \mu\text{mol photons}$

m⁻² s⁻¹, respectively. The lowest growth rate (0.016 day⁻¹) determined in this study was observed for both P2H7 (Greenland) and F5 (Chukchi Sea) grown under constant light (24h:0h L:D) and at the highest irradiance level (~366 μmol photons m⁻² s⁻¹). The next lowest growth rate was exhibited by isolate E516 (Greenland) under the aforementioned experimental conditions.

Although irradiance level had significant effects on growth, these effects varied according to photoperiod interval and among regions. Under the shortest light-period interval (16h:8h L:D), growth rates were generally lowest at the lowest irradiance level, and were significantly higher at moderate and higher irradiance levels (Fig. 7). However, under extended light-period treatments an inverse pattern was observed, whereby growth rates were highest under low and moderate light levels; apparent photoinhibition was most pronounced under constant light treatment (24h:0h L:D) (Fig. 7, Supplementary Fig. S1). Comparison of the combined growth response dataset among regions (data pooled from each irradiance level) showed that regardless of irradiance level, growth rates of isolates from the Chukchi Sea and GOM were significantly higher under the shortest irradiance interval (16h:8h L:D) than under an extended interval (20h:4h L:D) or constant light regime (24h:0h L:D) ($p<0.05$; Welch's ANOVA; Wilcoxon multiple comparisons). Growth rates of isolates from Greenland grown under the constant light were significantly lower than those grown under the other photoperiod intervals ($p<0.05$; Welch's ANOVA; Wilcoxon multiple comparisons). In contrast, no statistically significant differences in growth rates were observed among the different photoperiod intervals for the Iceland isolates.

Discussion

This is the first effort, to our knowledge, to investigate and compare the abundance and distribution of *Alexandrium fundyense* cysts in bottom sediments of coastal Greenland and Iceland, areas which have been recently impacted by toxic *Alexandrium* blooms. Our surveys of western Greenland and Iceland documented low to moderate *A. fundyense* cyst abundances (~20 to 400 cysts cm⁻³) throughout these regions; notably, cysts were observed at nearly all of stations surveyed, but sediments from Iceland contained substantially more cysts compared with samples from Greenland.

Based on the analysis of field data collected during the cruise and results of physiology experiments examining growth responses of Arctic *Alexandrium* isolates, we hypothesize that light availability and temperature regimes in the water column in fjords and coastal areas of Greenland are largely unfavorable for germling survival during transit from bottom waters to the surface, even during summer. However, many areas in Iceland could support germling survival and vegetative cell growth, which indicates that cyst deposits in Iceland may indeed be functioning for in situ bloom initiation. We note that the survival and excystment challenges faced by *Alexandrium* cysts in the deep fjords and cold waters of the Arctic are common to many other cyst-forming dinoflagellate species, and to spore-forming diatoms as well, so there is broad ecological relevance to our findings.

Cyst distribution

The accumulation rate and total abundance of cysts at a particular location reflects the net balance between deposition versus advective and germination losses, and is thus affected by bathymetric and hydrographic characteristics and processes

that determine bloom and/or cyst retention, as well as external and biological controls of cyst germination and bloom initiation. *Alexandrium* spp. cyst densities ranging from hundreds to thousands of cysts per cubic centimeter of surface sediments have been reported from areas around the world impacted by annual *Alexandrium* blooms and PSP, including the Gulf of Maine (GOM) of Canada and the USA, Puget Sound (USA) in the northeast Pacific, several coastal regions of Japan, and the western Mediterranean (Thau Lagoon, France). In the northwestern Atlantic, densities as high as 2000 cysts cm⁻³ and 6700 cysts cm⁻³ were reported from the Bay of Fundy and GOM, respectively (Anderson et al. 2014), and abundances >12,000 cysts cm⁻³ were observed in the Puget Sound region, in an area known to be a hot spot for PSP toxins in shellfish (Horner et al. 2011). Notably, extraordinarily high *A. fundyense* cyst densities (>10,000 cysts cm⁻³) in the Chukchi Sea were recently reported by Natsuike et al. (2013). However, in contrast with the aforementioned regions in which high cyst densities were generally associated with massive seasonal blooms, cyst concentrations in the Chukchi Sea sediments may well reflect the deposition of cysts year after year during a series of smaller blooms over time (rather than cyst deposition following a major bloom), with little or no germination losses, leading to the high abundances observed. As an example of the magnitude of cyst deposition following a major bloom, McGillicuddy et al. (2014) documented a red-water *A. fundyense* bloom in the GOM (cell densities in excess of 3×10^6 cells L⁻¹) that deposited only 10% as many cysts as observed in Chukchi Sea sediments (Natsuike et al. 2013).

With the exception of the two stations near Cape Farewell, *A. fundyense* cysts were found in all Greenland samples collected during the cruise. Cyst accumulations in Greenland sediments were generally low, and the maximum abundance of 37 cysts cm⁻³ was observed in Uummannaq Fjord (Fig. 3). In contrast, much higher cyst

abundances were observed in sediments from Iceland, ranging from 15–408 cysts cm⁻³. Although cyst accumulations in Greenland and Iceland were low to moderate compared with areas impacted by large-scale, annual *Alexandrium* blooms (e.g., Anderson et al. 2014), the concentrations we found are well within the range reported from areas with seasonal *Alexandrium* blooms and recurrent PSP toxin accumulation in shellfish. One example of relatively low cyst abundance levels leading to *Alexandrium* blooms comes from the Nauset estuary in Massachusetts, USA, where cyst densities of 150–418 cysts cm⁻³ have been associated with blooms and recurrent PSP-related shellfish harvesting closures in several of the embayments within that system (Crespo et al. 2011). Likewise, in a lagoon-wide survey of the Thau Lagoon in France, which is impacted annually by PSP toxin contamination in shellfish, the mean density of *Alexandrium* cysts was relatively low (<20 cysts g⁻¹ dry sediment [DS]), with the highest density (~440 cysts g⁻¹ DS) recorded at one location within the system where dense blooms were previously observed (Genovesi et al. 2013). Using the relationship between cyst abundance normalized to sediment dry weight versus sediment volume determined for *A. fundyense* in the GOM (Anderson et al. 2014), these Thau Lagoon values equate to 34–185 cysts cm⁻³. The GOM relationship may not be entirely appropriate for Thau Lagoon because of differences in sediment consistency and granularity, but is considered suitable for a rough approximation of cyst cm⁻³ levels.

Sediment structure

Although cyst distributions are frequently heterogeneous and sites-specific, prior investigations seeking to better define the physical dynamics underlying the occurrence of cyst seedbeds have identified several important characteristics common

to many important cyst accumulation zones. First and perhaps most importantly, higher cyst densities have been reported from protected or enclosed areas such as fjords, embayments, and harbors (Anderson 1997, Godhe & McQuoid 2003, Crespo et al. 2011), which serve to entrain blooms and promote local cyst deposition. Cyst abundance is also positively correlated with the proportion of finer grains and levels of total organic carbon (TOC) in sediments (Horner et al. 2011, Genovesi et al. 2013, Anderson et al. 2014). Finally, higher abundances have been linked with higher summer surface water temperature, which serves to stimulate dinoflagellate growth and promote the vertical stratification of the water column (Godhe & McQuoid 2003), leading to both a higher potential inoculum and reduced advective loss of cells within the system.

Dale (1976) first proposed that cysts tend to behave as fine silt particles in sediment dynamics, and as such, increase in abundance as the proportional abundance of finer sediment increases (often at depth). This hypothesis is supported by reports of higher *Alexandrium* spp. cyst accumulations in finer sediments or mud compared with sandy areas (Nehring 1994, Gayoso 2001, Yamaguchi et al. 2002, Anderson et al. 2005), and by subsequent investigations of the correlation between cyst densities and sediment characteristics. In surveys of Puget Sound, Washington, USA, Horner et al. (2011) observed a positive correlation between cyst abundance and the percentage of clay and silt in sediments during small scale surveys in Quartermaster Harbor, located in the south basin of Puget Sound. This pattern was not observed, however, in the large scale, sound-wide surveys, potentially due to variable and site-specific physical forcing conditions within the system (Moore et al. 2008, Horner et al. 2011).

Genovesi et al. (2013) documented a significant correlation between *Alexandrium* cyst densities and the *F*_{20–50} sediment fraction in the Thau Lagoon,

along the French Mediterranean coast, and reported that the cysts in this ecosystem effectively behaved like 20 to 50 μm particles (the approximate size range of *Alexandrium* cysts). With the exception of the two stations sampled near Cape Farewell, sediments in Greenland and Iceland were characterized by a high proportion of finer grained sediment fractions ($<63\ \mu\text{m}$) (Fig. 5), and would thus favor cyst accumulations in these areas. The most apparent difference among locations was the higher proportion of fine silt ($<2\ \mu\text{m}$) in the majority of samples collected from Greenland (Uummannaq and Disko Bay) compared with those from Iceland. These very fine particles, also referred to as glacial flour, are transported to the estuary with the melt water (Lund-Hansen et al. 2010). A second apparent difference was the higher proportion of coarser, sandy sediments ($>63\ \mu\text{m}$) in samples from St 523 and 524. Both stations are located near Cape Farewell on the southern coast of Greenland, and are typically exposed to open sea conditions with higher turbulent energy, thus preventing the accumulation of finer sediments. Notably, these were the only samples in which *Alexandrium* cysts were absent. Regional differences were also evident in the PCA of data on the sediment structure, cyst densities, and sampling depth (Fig. 6). This analysis identified at least three major clusters, grouped according to geographical region. One exception was St. 516 (Disko Bay), which clustered with stations sampled from Iceland in the PCA plot (Fig. 6), and was characterized by a lower proportion of sediment fractions $<6.3\ \mu\text{m}$ and a higher proportion of fractions $>63\ \mu\text{m}$ compared with the other samples from Uummannaq, Disko Bay, and the Vaigat (Fig. 5). This analysis also linked depth with the finest sediment fractions ($<6.3\ \mu\text{m}$), whereas cyst densities were linked with the intermediate ($F_{6.3-63}$) and coarsest size fraction ($F_{630-2000}$), the latter of which was only found in samples collected from St 516 and St 530 (Arnarfjörður).

Water temperature, depth, and bio-optical parameters

Cyst germination, and subsequent germling cell survival and growth, are highly dependent on light availability and temperature (Anderson 1980, Rengefors & Anderson 1998, Kremp & Anderson 2000, Vahtera et al. 2014); thus, the striking differences in these water column characteristics observed among sampling sites suggest that bloom development might only occur at certain locations within the study area. Temperatures throughout much of the water column in Greenland were generally $<4^{\circ}\text{C}$, with the maximum temperature of $\sim 12^{\circ}\text{C}$ only detected in surface waters at two locations (Fig. 2). Previous temperature measurements from Disko Bay during the summer months (June-August) ranged from $\sim 4\text{--}7^{\circ}\text{C}$ (Madsen et al. 2001, Heide-Jørgensen et al. 2007), thus the frequency and extent of the warmer surface waters we documented ($>10^{\circ}\text{C}$) is yet unknown. In contrast, water column temperatures measured in Iceland were much higher, ranging from 2 to 12°C in Arnarfjörður and 6 to 15°C in Breiðafjörður; with the exception of the deepest areas of Arnarfjörður ($>60\text{ m}$), water temperatures were generally $>8^{\circ}\text{C}$. Notably, Burrell et al. (2013) observed high cell concentrations ($>10,000\text{ cells L}^{-1}$) of *Alexandrium* spp., which they tentatively designated as *A. tamarense* along with small numbers of *A. ostenfeldii*, in Breiðafjörður and in Eyjafjörður (northern Iceland) in 2009. Low to moderate *Alexandrium* spp. cell concentrations were also observed in Breiðafjörður and Eyjafjörður from 2005–2008 (Gudfinnsson et al. 2010, Burrell et al. 2013), indicating that blooms may be recurrent at these locations.

Differences in bio-optical properties affecting light availability and quality over depth are important determinants of photosynthetically driven growth potential among *Alexandrium* populations at various locations. Based on the results of our

laboratory experiments, variation in day length expected in the study region would be most likely to promote growth in August, during which highest seawater temperatures would also be expected. Day length during summer months (July-August) in western Greenland (Disko Bay) and Iceland ranges from >20 hours during much of July, to between ~15-20 hours in August. In our laboratory experiments, highest growth rates were measured at irradiance levels of $\sim 150 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ or greater under the 16 h photoperiod interval. However, the comparatively high growth rates were also observed at low light levels under extended light-period treatments, indicating that longer photoperiod intervals may also be suitable for growth.

The 1% depth of PAR derived from our field data was interpreted to indicate the lowest depth of sufficient light for positive *Alexandrium* cell growth in the study region. In general, 1% PAR was surface bound to the upper water column, and was restricted to the top 50 m and 35 m for Greenland and Iceland, respectively (Table 1). Considering water depth, we inferred that the distance to be covered by vertically motile germling cells from bottom water to reach sufficient light for positive growth ranges from <70 m for most Iceland fjord stations to 500 m for St 515, the deepest Greenland station in this dataset. Assuming an average swimming speed of 10 m day⁻¹ (Eppley et al. 1968, Bauerfeind et al. 1986, Kamykowski et al. 1992) cells germinating at depths of 70 m would require approximately seven days to reach the surface, whereas cells germinating at 500 m would require 50 days. These transit times may be shorter, however, during upwelling conditions, which could rapidly transport cells to surface waters. The ability of germling cells to survive vertical transit to the euphotic zone in these areas will determine the potential for bloom initiation at these locations (see below).

Cyst viability and germling cell survival

Whether or not *A. fundyense* cysts in the Arctic and sub-Arctic (Greenland, Iceland, Chukchi Sea) are able to germinate, and corresponding vegetative cells to transit to the euphotic zone, under the particular temperature and light conditions present (i.e., very cold, deep, and dark waters) to initiate in situ blooms remains unknown. If not, the observed cyst deposits could represent end points or terminal deposits, with the bloom populations that ultimately produce those cysts originating from subarctic systems in the south through transport by coastal currents. Alternatively, cyst deposits from nearby shallow areas could serve as an initiation site for local blooms that lead to deposits in the deeper fjord sections.

To our knowledge, the potential for cyst germination and cell growth of *A. fundyense* (or any cyst-forming dinoflagellate) from Arctic and subarctic regions have not been studied. At high latitudes, cyst behavior and germling survival and growth at low temperatures and under an extended light-period interval in summer are of fundamental importance to bloom development and life cycle completion. Low temperatures can maintain cyst quiescence for extended periods (months, years, even decades) after cyst deposition and where germination is possible will also regulate the rate of excystment. Following excystment, the germling cell must survive the transit to surface waters, which is influenced by distance travelled in the dark (depth), availability of temperature and light, and cellular energy reserves (Vahtera et al. 2014). For the *A. fundyense* strains tested thus far from temperate waters, cyst germination either did not occur or proceeded at extremely low rates at temperatures between 0 and 4 °C (Anderson & Morel 1979, Anderson 1980). Based on the CTD measurements collected during our surveys, bottom temperatures in Greenland are expected to be in this range or lower during much of the potential *Alexandrium* bloom

season (Fig. 2). Anderson et al. (2005) showed that at 2 °C, *A. fundyense* cysts from the GOM required up to two months of incubation to reach 50% germination, whereas at 8 °C, this only took one to two weeks. At the low rates expected for cold Arctic waters, the bloom inoculum from excystment would be very gradual and slow, and might therefore introduce cells into the water too late in the season for successful bloom formation and new cyst deposition.

Following excystment, temperature and light are both important limiting factors that determine germling survival and vegetative cell growth. For many *A. fundyense* strains, including the few isolates examined from Greenland, Iceland, and the Chukchi Sea, a temperature range for survival and growth of 2 to 24 °C has been observed, with rates that are <25% of maxima at 6 °C or less (Watras et al. 1982; Anderson and Rengefors 2006; D.M. Anderson unpub. data). Although they were collected from Arctic and sub-arctic locations, the isolates we examined did not appear to be physiologically adapted for growth and survival in the extremely cold bottom water temperatures in the Arctic. Instead, their growth responses to temperature were similar to those of temperate isolates, with the maximum growth rate for all isolates found between 16–18 °C. These data will be published separately, along with a detailed analysis of the toxin contents of these isolates (Tillmann et al. 2016; D.M. Anderson unpub. data). The CTD temperature profiles collected during the cruise indicated that summer water temperatures in the Uummannaq/Disko Bay region only ranged from 4 to 8 °C, well below the temperature range for optimal growth, and bottom temperatures were much lower (Fig. 2).

Furthermore, the depth from which sediments were collected suggests that the survival of germinated cysts would be low at many of the locations surveyed in Greenland. Laboratory experiments examining the effects of dark treatment on cyst

germination and survival estimated that <50% of germinated cells would survive a 70 m transit from bottom sediments to the surface, and only 20% could survive a 200 m ascent in the dark (Vahtera et al. 2014). In Iceland, the estimated distances germlings would have to travel from germination depth to reach the 1% PAR depth ranged from 13 to 167 m; however, these estimated distances are substantially greater in Greenland, where the travel distance from germination depth to 1% PAR ranged from 68 to 503 m. Using the equations derived by Vahtera et al. (2014) describing the depth-related mortality rate, and assuming an initial survival time of one day, the proportions of cells estimated to survive the transit from the germination depth to 1% PAR ranged from 14 to 26% in Greenland, and 18 to 82% in Iceland.

Based on these field and experimental data, it is likely that *Alexandrium* cells and associated toxins in shellfish from Greenland are primarily derived from advected *Alexandrium* populations. There may also be certain shallow, nearshore areas, however, not explored in this study, that could provide favorable habitat for cyst germination and germling survival. Conditions in the bay and fjords we surveyed in western Iceland are suitable for germling survival and vegetative cell growth, and therefore may support indigenous, self-seeding blooms. The potential for *Alexandrium* bloom initiation in Greenland and other Arctic areas may be enhanced in the future, as Arctic Ocean bottom temperatures are projected to increase at a rate of 1 to 5 °C per 100 years, with a higher rate in nearshore regions (Bjastoch et al. 2011). This will clearly have an impact on the germination and survival rate of *Alexandrium*, but also will affect the distribution and bloom timing of many other meroplanktonic phytoplankton species.

Conclusions

610 Our field investigation documented low to moderate densities of *Alexandrium*
611 cysts in most areas surveyed in Greenland and Iceland, with highest densities
612 observed in western Iceland. We know that *A. fundyense* strains disperse readily and
613 are highly adaptable to new regions due to their ability to form cysts, overwinter, and
614 germinate to initiate blooms. Based on data collected on the temperature and light
615 availability (as influenced by water depth), we hypothesize that blooms in Greenland
616 are primarily derived from advected *Alexandrium* populations, as extremely low
617 bottom temperatures and travel distance from germination depth to the euphotic zone
618 would preclude in situ bloom initiation at most of the locations we surveyed.
619 Alternatively, cyst deposits from nearby shallow areas could serve as an initiation of
620 local blooms that lead to deposits in the deeper fjord sections. We further hypothesize
621 that in contrast with the situation in Greenland, the bays and fjords in Iceland provide
622 favorable habitat for germling cell survival and growth, and therefore may support
623 indigenous, self-seeding blooms.

624 The potential for *Alexandrium* blooms in Greenland and other Arctic areas
625 may change, as projected increases in water temperatures could expand habitat
626 suitable for *Alexandrium* germling survival and cell growth, particularly at nearshore
627 locations. The human health and ecosystem impacts of this potential expansion will
628 be significant, as marine bioresources are extremely important to the economies of
629 both Greenland and Iceland. Additional studies are needed to examine the physiology
630 of *Alexandrium* cysts and cells from the Arctic, particularly with regard to the
631 potential for cyst germination under ambient conditions in the region. These data will
632 help to further characterize processes that determine the distribution of endemic
633 versus introduced populations of *Alexandrium* and other toxin-producing

phytoplankton in the Arctic, and will be useful for understanding the potential for dispersal in the region under warmer conditions.

ACKNOWLEDGEMENTS

Funding for this study was provided by the James M. and Ruth P. Clark Arctic Research Initiative to Anderson and Richlen, and for the ARCHEMHAB expedition via the Helmholtz Institute initiative Earth and Environment under the PACES Program Topic 2 Coast (Workpackage 3) of the Alfred Wegener Institute. The research is part of the SCOR/IOC GEOHAB Core Research Project on HABs in Fjords and Coastal Embayments, Additional support was provided by the Woods Hole Center for Oceans and Human Health through National Science Foundation (NSF) Grant OCE-1314642 and National Institute of Environmental Health Sciences (NIEHS) Grant 1-P01-ES021923-01. We are grateful to Daniela Voß, Daniela Meier, and Rohan Henkel for their assistance during the cruise, and for helping to prepare the figures. We also thank Prof. Haifeng Gu for providing sediments from the Chukchi Sea, and Kerry Norton, Dave Kulis, John Brinckerhoff, Bruce Keafer, Lauren Henry, Hovey Clifford, and Judy Kleindinst for logistical and laboratory support and assistance. Finally, we acknowledge the generous support and assistance provided by Captain Klaus Bergman and crew of the Maria S. Merian throughout the cruise.

659 References

- 660 Anderson DM (1980) Effects of temperature conditioning on development and germination of
661 *Gonyaulax tamarens* (Dinophyceae) hypnozygotes. *Journal of Phycology* 16:166-172
- 662 Anderson DM (1997) Bloom dynamics of toxic *Alexandrium* species in the northeastern U.S.
663 *Limnology and Oceanography* 42:1009-1022
- 664 Anderson DM, Fukuyo Y, Matsuoka K (2003) Cyst methodologies. In: Hallegraeff GM (ed) Manual on
665 harmful marine microalgae, Book 11. UNESCO, Monographs on Oceanographic
666 Methodology
- 667 Anderson DM, Keafer BA, Kleindinst JL, McGillicuddy Jr DJ, Martin JL, Norton K, Pilskaln CH,
668 Smith JL, Sherwood CR, Butman B (2014) *Alexandrium fundyense* cysts in the Gulf of
669 Maine: Long-term time series of abundance and distribution, and linkages to past and future
670 blooms. *Deep Sea Research Part II: Topical Studies in Oceanography* 103:6-26
- 671 Anderson DM, Morel FMM (1979) The seeding of two red tide blooms by the germination of benthic
672 *Gonyaulax tamarens* hypnocyts. *Estuarine and Coastal Marine Science* 8:279-293
- 673 Anderson DM, Stock CA, Keafer BA, Bronzino Nelson A, Thompson B, McGillicuddy JDJ, Keller M,
674 Matrai PA, Martin J (2005) *Alexandrium fundyense* cyst dynamics in the Gulf of Maine. *Deep*
675 *Sea Research Part II: Topical Studies in Oceanography* 52:2522-2542
- 676 Anderson DM, Taylor CD, Armbrust EV (1987) The effects of darkness and anaerobiosis on
677 dinoflagellate cyst germination. *Limnology and Oceanography* 32:340-351
- 678 Baggesen C, Moestrup Ø, Daugbjerg N, Krock B, Cembella AD, Madsen S (2012) Molecular
679 phylogeny and toxin profiles of *Alexandrium tamarens* (Lebour) Balech (Dinophyceae) from
680 the west coast of Greenland. *Harmful Algae* 19:108-116
- 681 Bauerfeind E, Elbrächter M, Steiner R, Thronsen J (1986) Application of Laser Doppler Spectroscopy
682 (LDS) in determining swimming velocities of motile phytoplankton. *Marine Biology* 93:323-
683 327
- 684 Biastoch A, Treude T, Rüpke LH, Riebesell U, Roth C, Burwicz EB, Park W, Latif M, Böning CW,
685 Madec G, Wallmann K (2011) Rising Arctic Ocean temperatures cause gas hydrate
686 destabilization and ocean acidification. *Geophysical Research Letters* 38:L08602
- 687 Burrell S, Gunnarsson T, Gunnarsson K, Clarke D, Turner AD (2013) First detection of paralytic
688 shellfish poisoning (PSP) toxins in Icelandic mussels (*Mytilus edulis*): Links to causative
689 phytoplankton species. *Food Control* 31:295-301
- 690 Cembella A, Zielinski O, Anderson D, Graeve M, Henkel R, John U, Kattner G, Koch B, Krock B,
691 Meier D, Richlen M, Tillmann U, Voß D (2016) ARCHEMHAB: Interactions and feedback
692 mechanisms between hydrography, geo-chemical signatures and microbial ecology, with a
693 focus on HAB species diversity, biogeography and dynamics Cruise Report MSM21/3, DFG-
694 Senatskommission für Ozeanographie, Bremen, Germany, 51 pages
- 695 Crespo BG, Keafer BA, Ralston DK, Lind H, Farber D, Anderson DM (2011) Dynamics of
696 *Alexandrium fundyense* blooms and shellfish toxicity in the Nauset Marsh System of Cape
697 Cod (Massachusetts, USA). *Harmful Algae* 12:26-38
- 698 Dale B (1976) Cyst formation, sedimentation, and preservation: Factors affecting dinoflagellate
699 assemblages in recent sediments from trondheimsfjord, Norway. *Review of Palaeobotany and*
700 *Palynology* 22:39-60
- 701 Eppley RW, Holm-Harisen O, Strickland JDH (1968) Some observations on the vertical migration of
702 dinoflagellates S. *Journal of Phycology* 4:333-340
- 703 Etheridge SM, Roesler CS (2005) Effects of temperature, irradiance, and salinity on photosynthesis,
704 growth rates, total toxicity, and toxin composition for *Alexandrium fundyense* isolates from
705 the Gulf of Maine and Bay of Fundy. *Deep Sea Research Part II: Topical Studies in*
706 *Oceanography* 52:2491-2500
- 707 Fraga S, Sampedro N, Larsen J, Moestrup Ø, Calado AJ (2015) Arguments against the proposal 2302
708 by John & al. to reject the name *Gonyaulax catenella* (*Alexandrium catenella*). *Taxon* 64:634-
709 635
- 710 Garaba SP, Zielinski O (2013) Comparison of remote sensing reflectance from above-water and in-
711 water measurements west of Greenland, Labrador Sea, Denmark Strait, and west of Iceland.
712 *Opt Express* 21:15938-15950, doi: 15910.11364/OE.15921.015938
- 713 Gayoso AM (2001) Observations on *Alexandrium tamarens* (Lebour) Balech and other dinoflagellate
714 populations in Golfo Nuevo, Patagonia (Argentina). *Journal of Plankton Research* 23:463-468
- 715 Genovesi B, Mouillot D, Laugier T, Fiandrino A, Laabir M, Vaquer A, Grzebyk D (2013) Influences of
716 sedimentation and hydrodynamics on the spatial distribution of *Alexandrium*

717 *catenella/tamarensis* resting cysts in a shellfish farming lagoon impacted by toxic blooms.
718 Harmful Algae 25:15-25

719 Godhe A, McQuoid MR (2003) Influence of benthic and pelagic environmental factors on the
720 distribution of dinoflagellate cysts in surface sediments along the Swedish west coast. Aquatic
721 Microbial Ecology 32:185-201

722 Gu H, Zeng N, Xie Z, Wang D, Wang W, Yang W (2013) Morphology, phylogeny, and toxicity of
723 Atama complex (Dinophyceae) from the Chukchi Sea. Polar Biol 36:427-436

724 Gudfinnsson HG, Eydal A, Gunnarsson K, Gudmundsson K, Valsdóttir K (2010) Monitoring of toxic
725 phytoplankton in three Icelandic fjords. ICES theme session N

726 Guillard RR, Ryther JH (1962) Studies of marine diatoms. I. Cyclotella nana Husdedt and Detonula
727 confervacea Gran. . Can J Microbiol 8:229-239

728 Guillard RRL (1973) Division rates. In: Stein JR (ed) Handbook of phycological methods. Cambridge
729 Univ. Press

730 Heide-Jørgensen MP, Laidre KL, Logsdon ML, Nielsen TG (2007) Springtime coupling between
731 chlorophyll a, sea ice and sea surface temperature in Disko Bay, West Greenland. Progress in
732 Oceanography 73:79-95

733 Holinde L, Zielinski O (2015) Bio-optical characterization and light availability parameterization in
734 Uummannaq Fjord and Vaigat-Disko Bay (West Greenland). Ocean Science 12:117-128,
735 doi:10.5194/os-5112-5117-2016.

736 Horner RA, Greengrove CL, Davies-Vollum KS, Gawel JE, Postel JR, Cox AM (2011) Spatial
737 distribution of benthic cysts of *Alexandrium catenella* in surface sediments of Puget Sound,
738 Washington, USA. Harmful Algae 11:96-105

739 John U, Litaker RW, Montresor M, Murray S, Brosnahan ML, Anderson DM (2014a) Formal Revision
740 of the *Alexandrium tamarensis* Species Complex (Dinophyceae) Taxonomy: The Introduction
741 of Five Species with Emphasis on Molecular-based (rDNA) Classification. Protist 165:779-
742 804

743 John U, Litaker W, Montresor M, Murray S, Brosnahan ML, Anderson DM (2014b) (2302) Proposal to
744 reject the name *Gonyaulax catenella* (*Alexandrium catenella*)(Dinophyceae). Taxon 63:932

745 Kamykowski D, Reed RE, Kirkpatrick GJ (1992) Comparison of sinking velocity, swimming velocity,
746 rotation and path characteristics among six marine dinoflagellate species. Marine Biology
747 113:319-328

748 Kremp A, Anderson DM (2000) Factors regulating germination of resting cysts of the spring bloom
749 dinoflagellate *Scrippsiella hangoei* from the northern Baltic Sea. Journal of Plankton Research
750 22:1311-1327

751 Lilly EL, Halanich KM, Anderson DM (2007) Species boundaries and global biogeography of the
752 *Alexandrium tamarensis* complex (Dinophyceae). Journal of Phycology 43:1329-1338

753 Lund-Hansen LC, Andersen TJ, Nielsen MH, Pejrup M (2010) Suspended matter, Chl-a, CDOM, grain
754 sizes, and optical properties in the Arctic fjord-type estuary, Kangerlussuaq, West Greenland
755 during summer. Estuaries and coasts 33:1442-1451

756 Madsen SD, Nielsen TG, Hansen BW (2001) Annual population development and production by
757 *Calanus finmarchicus*, *C. glacialis* and *C. hyperboreus* in Disko Bay, western Greenland.
758 Marine Biology 139:75-93

759 McGillicuddy Jr DJ, Brosnahan ML, Couture DA, He R, Keafer BA, Manning JP, Martin JL, Pilskaln
760 CH, Townsend DW, Anderson DM (2014) A red tide of *Alexandrium fundyense* in the Gulf of
761 Maine. Deep Sea Research Part II: Topical Studies in Oceanography 103:174-184

762 Moore SK, Mantua NJ, Kellogg JP, Newton JA (2008) Local and large-scale climate forcing of Puget
763 Sound oceanographic properties on seasonal to interdecadal timescales. Limnol Oceanogr
764 53:1746-1758

765 Natsuike M, Nagai S, Matsuno K, Saito R, Tsukazaki C, Yamaguchi A, Imai I (2013) Abundance and
766 distribution of toxic *Alexandrium tamarensis* resting cysts in the sediments of the Chukchi Sea
767 and the eastern Bering Sea. Harmful Algae 27:52-59

768 Nehring S (1994) Spatial distribution of dinoflagellate resting cysts in recent sediments of Kiel Bight,
769 Germany (Baltic Sea). Ophelia 39:137-158

770 Okolodkov YB (2005) The global distributional patterns of toxic, bloom dinoflagellates recorded from
771 the Eurasian Arctic. Harmful Algae 4:351-369

772 Preston-Thomas H (1990) The International Temperature Scale of 1990 (ITS-90). Metrologia 27.1:3-10

773 Rengefors K, Anderson DM (1998) Environmental and endogenous regulation of cyst germination in
774 two freshwater dinoflagellates. Journal of Phycology 34:568-577

775 Scholin CA (1998) Morphological, genetic and biogeographic relationships of *Alexandrium tamarensis*,
776 *A. catenella* and *A. fundyense*. In: Anderson DM, Hallegraeff GM, Cembella AD (eds) The

- Physiological Ecology of Harmful Algal Blooms. NATO Advanced Study Institute Series, Springer-Verlag, Heidelberg
- Schwinghamer P, Anderson DM, Kulis DM (1991) Separation and concentration of living dinoflagellate resting cysts from marine sediments via density-gradient centrifugation. *Limnol & Oceanogr* 36:588-295
- Tillmann U, Krock B, Alpermann TJ, Cembella A (2016) Bioactive compounds of marine dinoflagellate isolates from western Greenland and their phylogenetic association within the genus *Alexandrium*. *Harmful Algae* 51:67-80
- UNESCO (1983) Algorithms for computation of fundamental properties of seawater UNESCO technical papers in marine science, 44
- Vahtera E, Crespo BG, McGillicuddy DJ, Olli K, Anderson DM (2014) *Alexandrium fundyense* cyst viability and germling survival in light vs. dark at a constant low temperature. *Deep Sea Research Part II: Topical Studies in Oceanography* 103:112-119
- Watras CJ, Chisholm SW, Anderson DM (1982) Regulation of growth in an estuarine clone of *Gonyaulax tamarens* Lebour: salinity-dependent temperature responses. *Journal of Experimental Marine Biology and Ecology* 62:25-37
- Wood AM, Everroad R, Wingard L (2005) Measuring growth rates in microalgal cultures. In: Anderson RA (ed) *Algal culturing techniques*. Elsevier Academic Press, Burlington, MA
- Yamaguchi M, Itakura S, Nagasaki K, Kotani Y (2002) Distribution and abundance of resting cysts of the toxic *Alexandrium* spp. (Dinophyceae) in sediments of the western Seto Inland Sea, Japan. *Fisheries Science* 68:1012-1019

FIGURE LEGENDS

Figure 1. Map showing sediment sampling locations in Greenland and Iceland. Intensive sampling was carried out in the Disko Bay region (DB) and western Iceland (WI).

Figure 2. CTD profiles over the section distance in kilometers for temperature (°C) within selected sections in Greenland (Disko Bay region, top panel) and Iceland (Arnarfjörður and Breiðafjörður; middle and bottom panels, respectively). Stations identified by grey vertical lines. Greenland: starting point was the innermost station in Uummannaq fjord (Perlerfiup Sermia glacier). Vaigat was entered at section distance 250 km and Disko Bay at section distance 450 km. Arnarfjörður: starting point was the innermost station of the fjord. Breiðafjörður: starting point was the innermost station of the fjord.

Figure 3. Abundance and distribution of *Alexandrium fundyense* resting cysts (cysts cm⁻³) in sediments collected from Greenland (DB=Disko Bay region) and Iceland (WI=West Iceland).

Figure 4. *Alexandrium fundyense* cyst abundance (cysts cm⁻³) in sediments collected from Greenland and Iceland versus sampling depth.

Figure 5. Proportion of each sediment class in samples collected from Greenland and Iceland. A total of seven granulometric fractions (µm) were quantified. Sampling

locations: Uq=Uummannaq; Vg=Vaigat; DB=Disko Bay; CF=Cape Farewell; Af= Arnarfjörður; Bf=Breiðafjörður.

Figure 6. Principal component analysis (PCA) of *Alexandrium* cyst abundance, sediment characteristics, and sampling depth of sediments collected from Greenland (Uummannaq, Vaigat, Disko Bay, Cape Farewell) and Iceland (Arnarfjörður, Breiðafjörður). Stations in Greenland and Iceland are delineated by solid and dashed lines, respectively. Symbols denote specific sampling locations.

Figure 7. Growth rates (μ [day⁻¹]) of *Alexandrium fundyense* isolates from Greenland (n=3), Iceland (n=3), the Chukchi Sea (n=3), and the Gulf of Maine (n=3) in response to irradiance. Data from three different photoperiod intervals (L:D) are shown: 16h:8h (solid circles), 20h:4h (open triangles), and 24h:0 (shaded squares).

Supplementary Figure S1. Growth rates (μ [day⁻¹]) of *Alexandrium fundyense* isolates from Greenland (n=3), Iceland (n=3), the Chukchi Sea (n=3), and the Gulf of Maine (n=3) in response to irradiance and photoperiod interval. Data from three different photoperiod intervals (L:D) are shown: 16h:8h (top row), 20h:4h (middle row), and 24h:0 (bottom row).

1 TABLES

2 **Table 1.** Light availability from radiometric profiles. 1% PAR is the 1% depth level (m) of photosynthetically active radiation PAR(z) (μmol
3 photons $\text{m}^{-2} \text{s}^{-1}$) with respect to surface PAR(0⁺). $\lambda_{\text{max1\%}}$ is the maximum wavelength observed at that depth. Water depth is the bottom depth of
4 the respective station. All values are the mean of two to three profiler casts at the specific stations. “*” denotes stations where *Alexandrium* cysts
5 were quantified.

Station	Location	Water depth (m)	1% PAR	PAR(0+)	Lambda_max1%
503*	Uummannaq, Greenland	402.0	38.0	557	496
504*		350.0	35.7	194	496
506*		143.9	32.8	189	496
514*	Disko Bay, Greenland	259.1	41.3	78	492
515*		543.7	40.3	1291	495
516*		169.5	33.2	354	499
517	West Greenland coast	112.6	41.5	166	496
521		114.7	36.3	1076	496
522		113.2	45.7	1201	495
527*	Arnarfjörður, Iceland	54.0	24.6	332	541
528*		106.2	28.5	461	536
529*		103.2	27.8	289	538

530*		81.0	27.1	287	539
531		46.2	33.4	81	498
532	Breiðafjörður, Iceland	47.0	34.4	131	497
533		53.2	29.7	475	502
534		69.2	20.3	760	565
535*		55.6	15.9	366	567
536		34.2	19.6	88	563
537*		126.6	20.3	676	564
538*		189.9	22.5	966	562

1

2

3

4

5

6

7

8

1

2 **Supplementary Table S1.** Details regarding isolates used to characterize growth

3 responses to light intensity and photoperiod interval.

Isolate	Origin	Year	Isolation method
E516	Disko Bay, Greenland	2012	Single cell from cyst germination
P2E6		2012	Single vegetative cell
P2H7		2012	
P3H8		2012	
E9	Arnarfjörður, Iceland	2012	Single cell from cyst germination
D3		2012	
B10	Breiðafjörður, Iceland	2012	Single cell from cyst germination
GOM H15	Gulf of Maine	2005	Single vegetative cell
GOM D2		2005	
GOM F14		2005	
F5	Chukchi	2013	Single cell from cyst germination
C7		2013	
E2		2013	

4

5

6

7

8

9

10

11

12

13

14

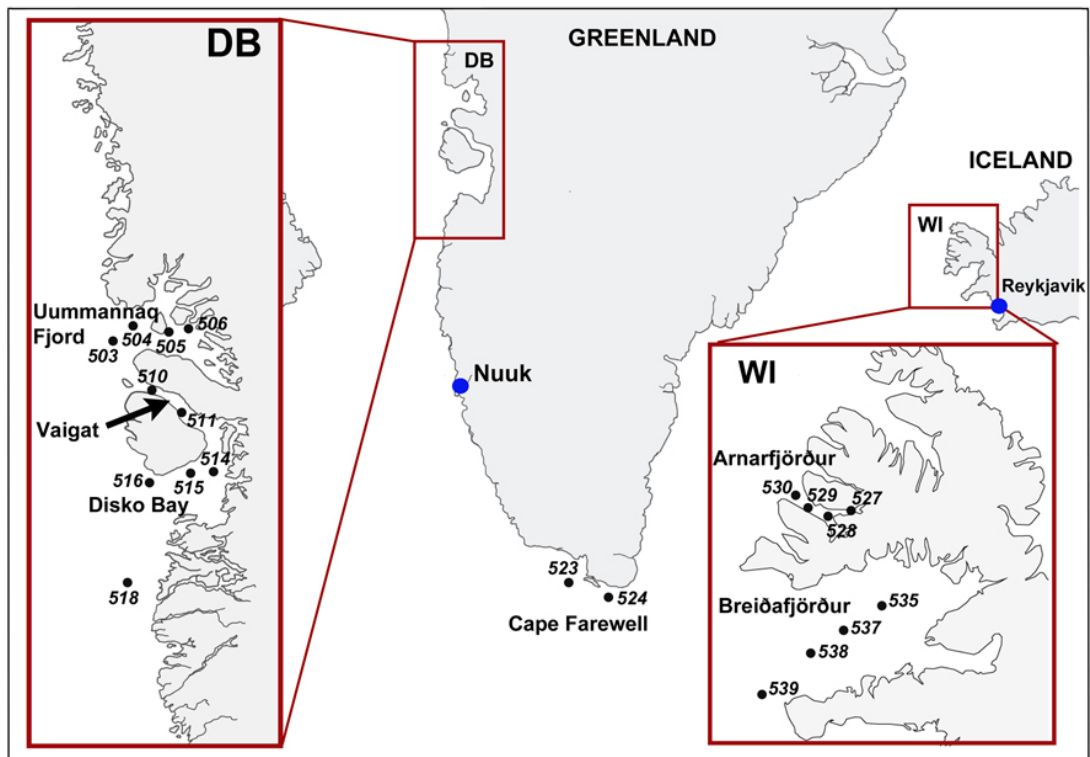
1

2 **FIGURES**

3

4 Fig. 1

5



6

7

8

9

10

11

12

13

14

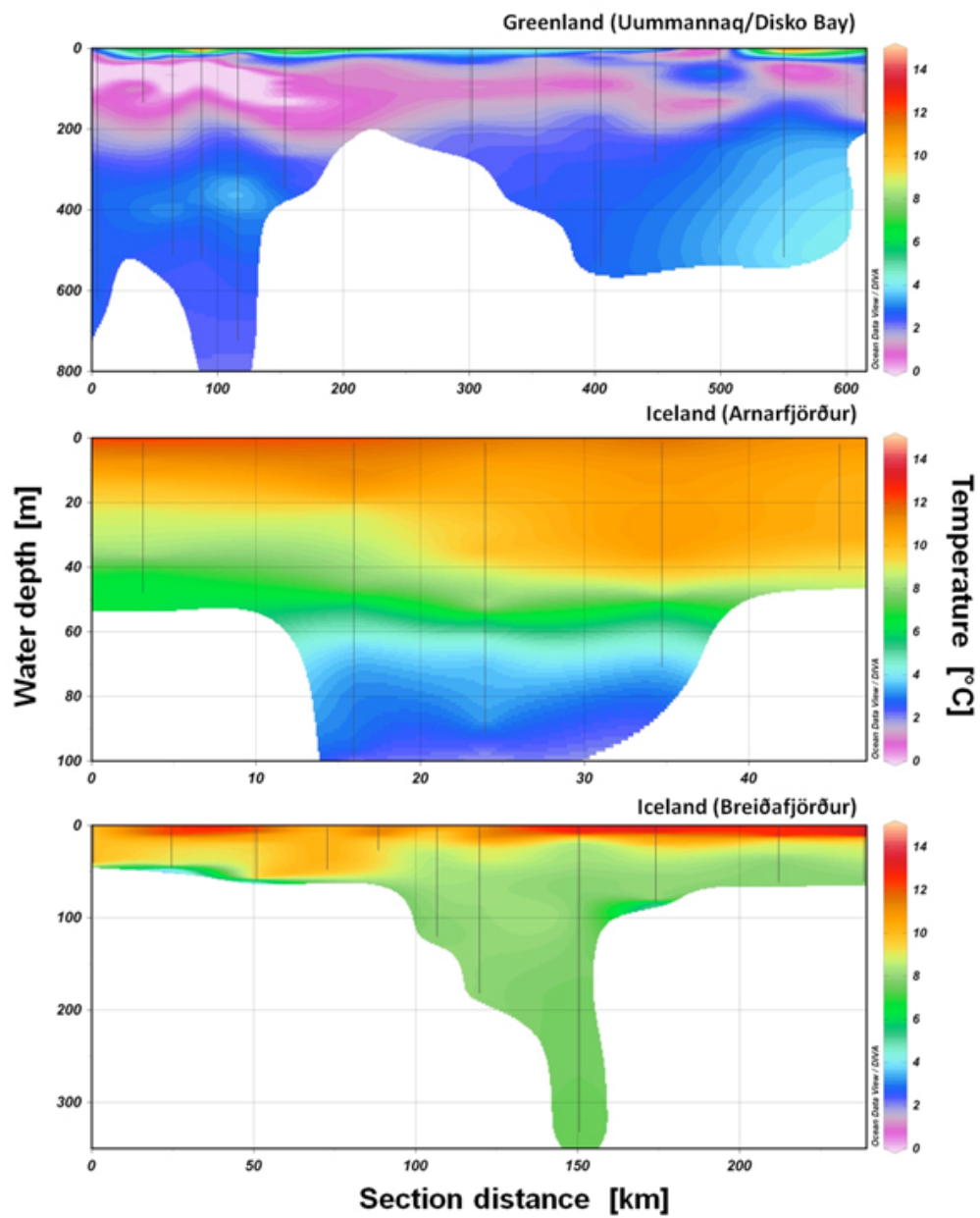
15

1

2

3 Fig. 2

4



5

6

7

8

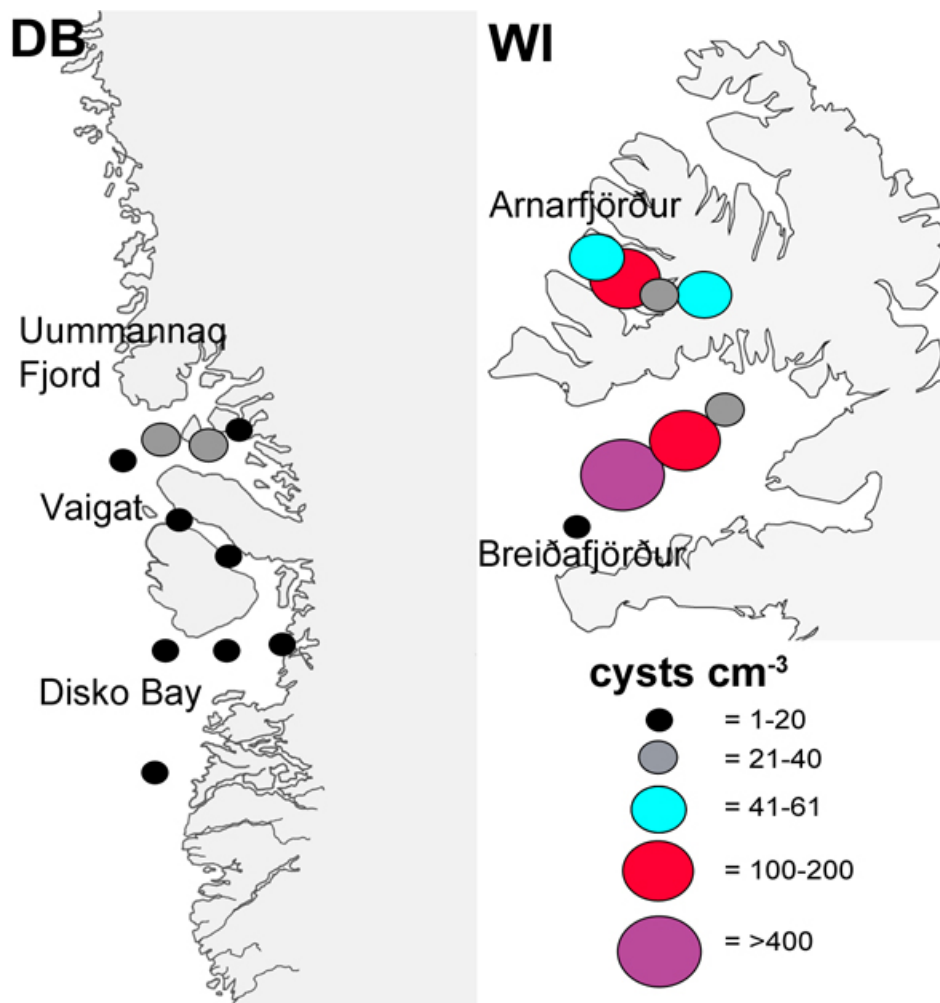
9

1

2

3 Fig. 3

4



5

6

7

8

9

10

11

12

Fig. 4

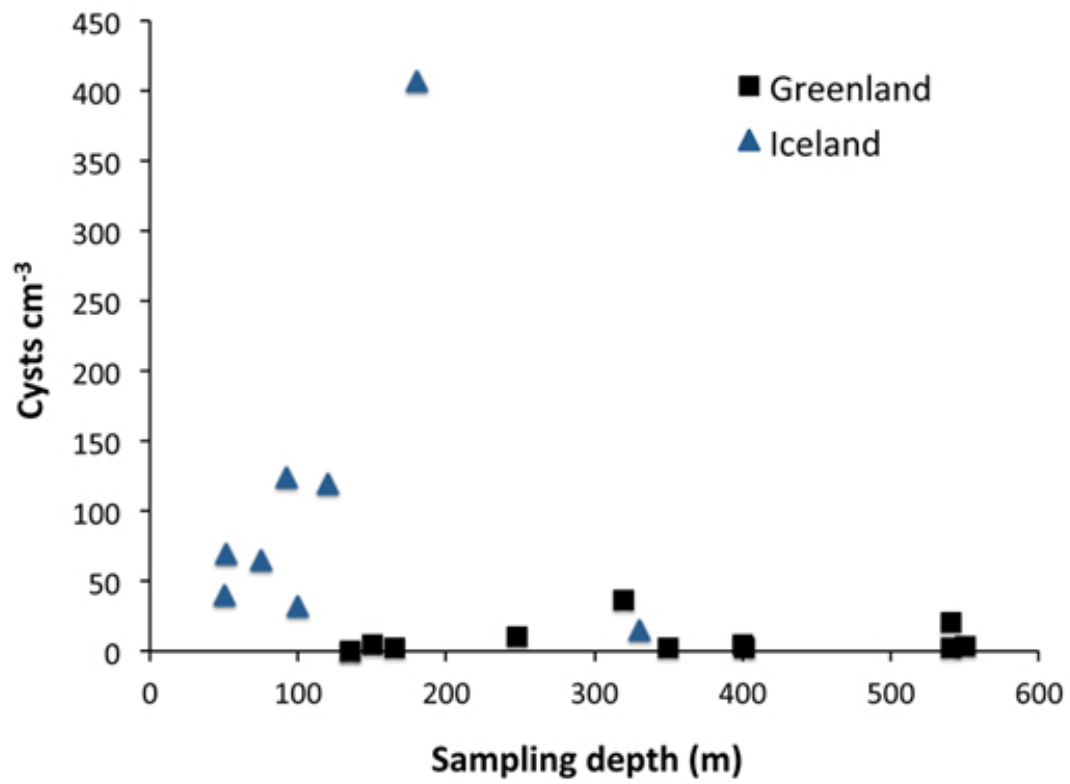


Fig. 5

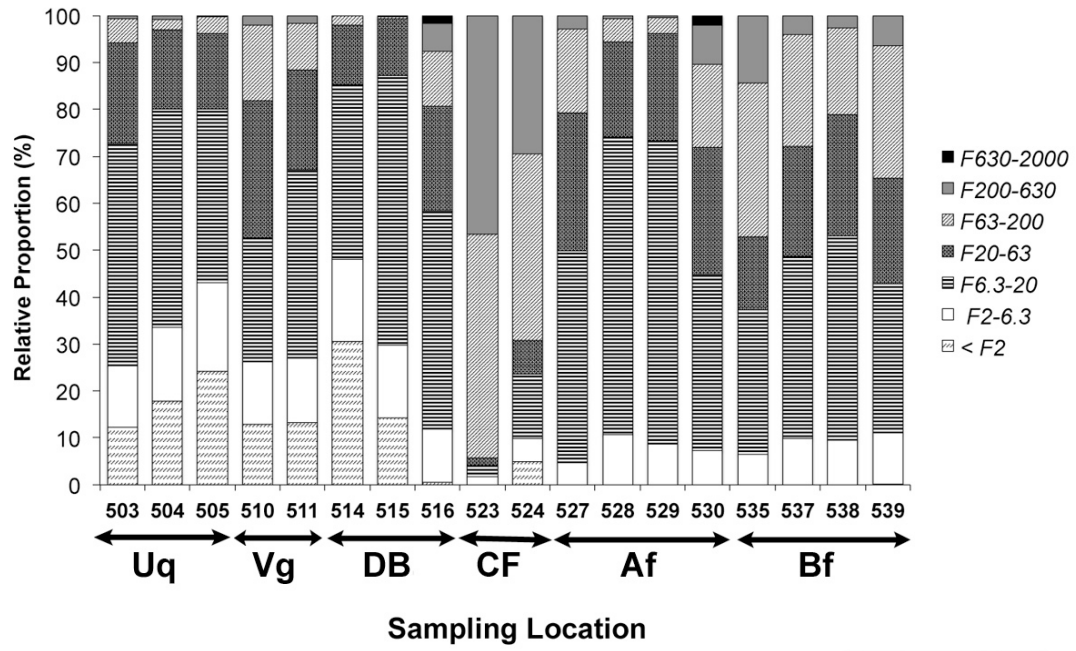
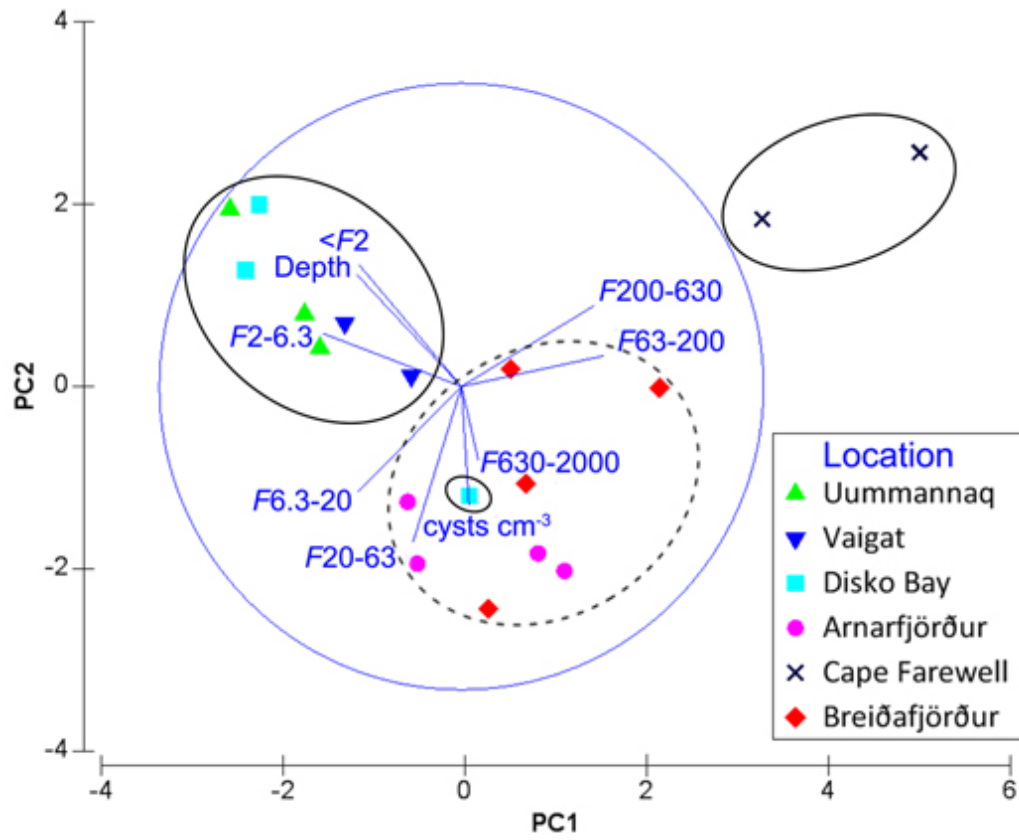


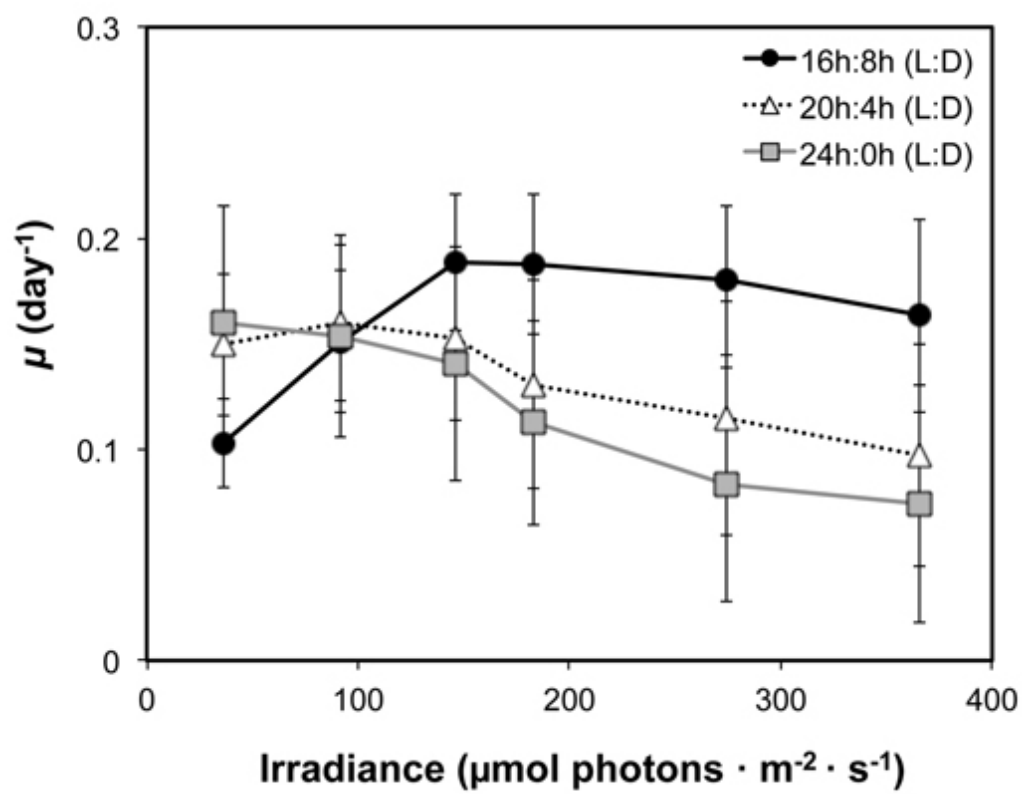
Fig. 6



1

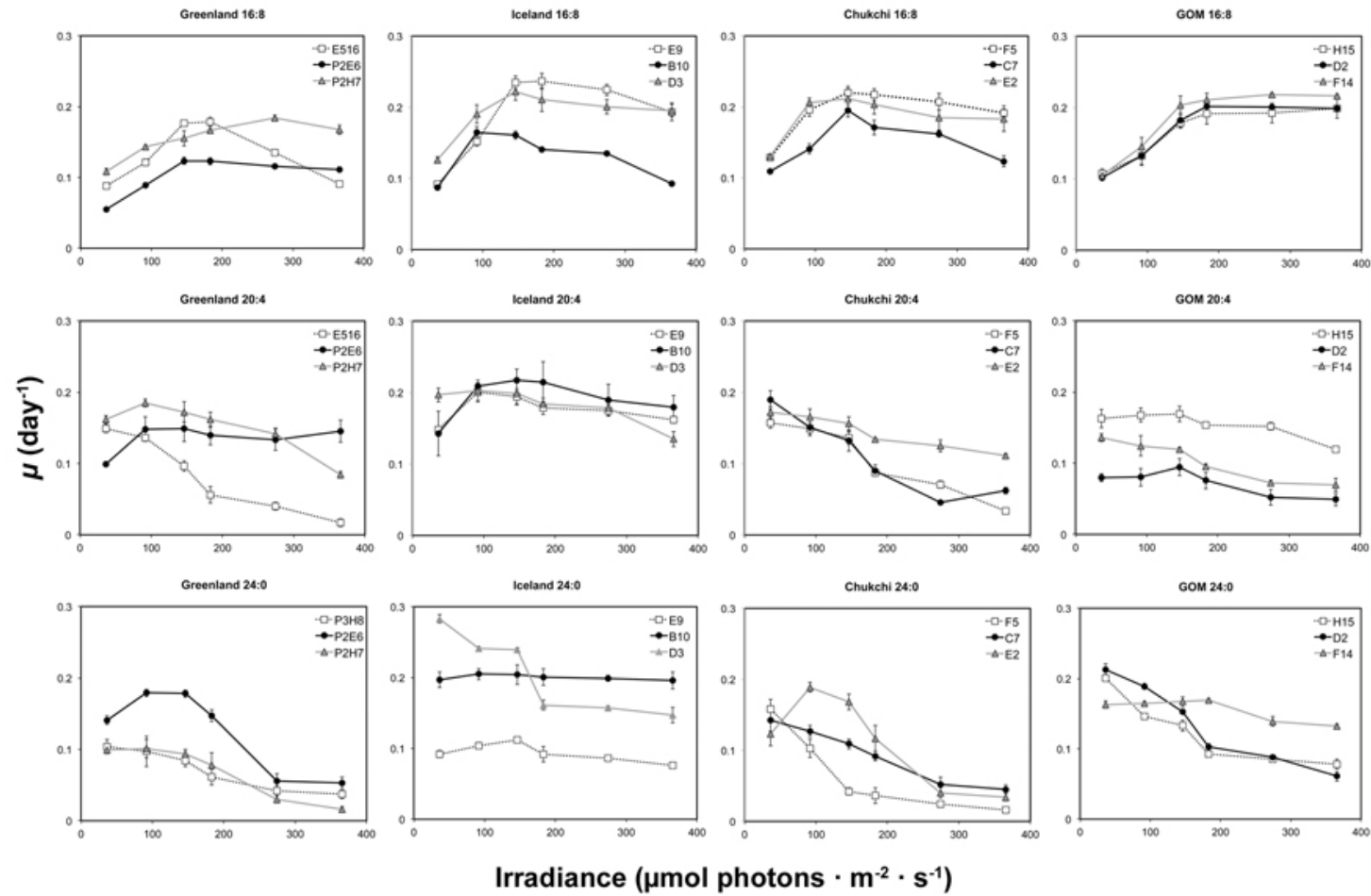
2 Fig. 7

3



4

1 Supplementary Figure S1.



2